Optimisation and performance evaluation of a semi-pilot scale iHydroMET for decentralised domestic wastewater treatment

Siddhant Sahoo

A dissertation submitted for the partial fulfilment of BS-MS dual degree in Science



Indian Institute of Science Education and Research Mohali May 2021

Certificate of Examination

This is to certify that the dissertation titled "Optimisation and performance evaluation of a semi-pilot scale iHydroMET for decentralised domestic wastewater treatment" submitted by Mr. Siddhant Sahoo (Reg. No. MS16138) for the partial fulfilment of BS-MS dual degree programme of the Institute, has been examined by the thesis committee duly appointed by the Institute. The committee finds the work done by the candidate satisfactory and recommends that the report be accepted.

Dr. Santosh B. Satbhai

Dr. Baerbel Sinha

Dr. Sunil A. Patil

(Supervisor)

Dated: May 21, 2021

Declaration

The work presented in this dissertation has been carried out by me under

the guidance of Dr. Sunil A. Patil at the Indian Institute of Science

Education and Research Mohali.

This work has not been submitted in part or in full for a degree, a diploma,

or a fellowship to any other university or institute. Whenever contributions

of others are involved, every effort is made to indicate this clearly, with due

acknowledgement of collaborative research and discussions. This thesis is

a bonafide record of original work done by me and all sources listed within

have been detailed in the bibliography.

Siddhant Sahoo

(Candidate)

Dated: May 21, 2021

In my capacity as the supervisor of the candidate's project work, I certify

that the above statements by the candidate are true to the best of my

knowledge.

Dr. Sunil A. Patil

(Supervisor)

iii

Acknowledgements

The following work is a result of sincere efforts and hard work that was not possible without the extensive support, constant encouragement, and expertise of different people who need a mention over here.

Firstly I would like to express my gratitude to Dr. Sunil Anil Patil for being such a wonderful guide and mentor. I wish to express my deep sense of appreciation and gratitude to him for the valuable guidance, motivation, and innovative inputs. Without his constant guidance and support, this dissertation would have been a tough nut to crack.

Besides my advisor, I would like to thank my thesis evaluation committee members, Dr. Baerbel Sinha and Dr. Santosh B. Satbhai, for their insightful comments and encouragement.

I would like to sincerely thank Mr. Ravi Kumar Yadav for his guidance, which helped me address the problems I faced during experiments and writing the thesis. Without him, I would not be in a place to complete this dissertation. He helped me immensely in carrying out the necessary experiments and taught me all the expertise needed for wastewater research since I joined the lab.

This acknowledgment would remain incomplete without expressing my gratitude to all EEMB family members for providing a helpful and friendly lab environment and helping me in lab work and beyond. I also wish to show my gratitude to the faculties and staff of the EES department for their continuous support and for availing lab facilities throughout this journey.

Now, I would like to thank Bhumija Gautam and Chetan Sadhotra for all the fun we had during this time, which helped me remain fresh and mentally active, and for the sleepless nights we were working together before deadlines.

I would also like to especially thank Broti Biswas and Puneeth Deraje for being by my side throughout and helping me. I am thankful to them for making my life so beautiful and memorable at IISER.

I would like to thank my colleague, Pragati Arora, for helping me with the bed matrices experiments.

I am thankful to the institute's Library for providing world-class library facilities, valuable e-resources, and round-the-clock and personalized services, and IISER Mohali for giving me the opportunity to explore the field of Science and KVPY for providing me the fellowship.

Lastly, these acknowledgements would be incomplete without expressing my thanks to the most important part of my life, my parents, Mr. Niranjan Sahoo and Mrs. Babita Sahoo. Thus, I express my profound gratitude towards them for giving me continuous motivation during my studies and supporting my decisions.

Siddhant Sahoo

List of Figures

Fig. 1. Schematic of different tested bed matrices in the reactor unit
Fig. 2. Schematic of the experimental setup for harvesting electricity in two cathode
positions: a) ¾ immersed in the effluent chamber and b) top of the reactor unit
Fig. 3. Schematic of the experimental setup used to evaluate nutrient removal performance
of the test plants: a) Vinca and b) Pothos
Fig. 4. Schematic of a single reactor unit configuration with the major components 16
Fig. 5. Schematic of the upgraded iHydroMET system
Fig. 6. Closed-circuit voltage in two cathode placement conditions for 6h
Fig. 7. Pictorial view showing the increment in shoot biomass of representative tested plants
grown for 1 month: a) Vinca; b) Pothos; c) Peppermint; d) Brahmi, and e) Lemongrass 22
Fig. 8. Removal of total nitrogen (TN) by the tested plants for 48h period
Fig. 9. Removal of total phosphorus (TP) by the tested plants for 48h period
Fig. 10. Removal efficiencies of various pollutants by iHydroMET at different HRTs 25
Fig. 11. Polarization (orange) and Power density (blue) curves of a) serially connected
MFCs, and b) parallelly connected MFCs of the iHydroMET system
Fig. 12. Closed-circuit potential curves for 6h at optimum resistance with the a) serially
connected MFCs, and b) parallelly connected MFCs of the iHdroMET system 30

List of Photographs

Photograph 1. a) Different materials tested as the bed matrix component	ents: a) cocopeat, b)
granular activated charcoal, and c) quartz sand	10
Photograph 2. Different test plants planted in domestic wastewater-fed	d reactor units 14

List of Tables

Table 1. COD, ammoniacal-N, and orthophosphate removal efficiencies (%) achieved by
different support bed material matrices in reactor units.	20
Γable 2. Consolidated data of wastewater treatment performance	parameters of
HydroMET at different HRTs	24

Abbreviations

BM Bed Matrix

BOD Biological Oxygen Demand

CCV Closed Circuit Voltage

COD Chemical Oxygen Demand

CP Cocopeat

CWs Constructed Wetlands

DS Domestic Sewage

GAC Granular Activated Charcoal

HRT Hydraulic Retention Time

MFC Microbial Fuel Cell

MLD Million Liters Per Day

MPN Most Probable Number

N Nitrogen

P Phosphorus

PAOs Phosphate Accumulating Organisms

PPCPs Pharmaceuticals and Personal Care Products

QS Quartz Sand

SDGs Sustainable Development Goals

STP Sewage Treatment Plants

TN Total Nitrogen

TP Total Phosphorus

UV Ultraviolet

Contents

Lis	st of Figures	vi
Lis	st of Photograph	vii
Lis	st of Tables	viii
At	obreviations	ix
At	ostract	1
1.	Introduction	3
	1.1. Wastewater: An environmental problem but promising solution to ad	ldress water
	stress	3
	1.2. Existing wastewater treatment infrastructure	4
	1.3. The era of decentralization	5
	1.4. Emerging decentralised technologies for wastewater treatment	6
	1.5. The idea of iHydroMET	7
2.	Objectives	8
3.	Materials and Methods	9
	3.1. Selection and optimisation of support bed matrix components	9
	3.1.1. Cocopeat	9
	3.1.2. Granular activated charcoal	10
	3.1.3. Quartz sand	10
	3.1.4. Methodology for bed matrix optimisation	10
	3.2. Cathode position optimisation	12
	3.3. Plant selection.	13
	3.4. Performance evaluation of the upgraded iHydroMET system with the	e optimised
	set of reactor components	15
	3.5. Analyses and calculations	17
	3.5.1. Wastewater parameters	17
	3.5.2. Electrochemical parameters	18
4.	Results and Discussion.	19
	4.1. Selection and optimisation of support bed matrix components	19
	4.2 Cathode position optimistion	20

4.3. Plant	Selection	21
4.4. Perfor	rmance evaluation of the upgraded iHydroMET with the optim	nised reactor
comp	onents	23
4.4.1.	Organics	25
4.4.2.	Total Phosphate	26
4.4.3.	Total Nitrogen	26
4.4.4.	pH	27
4.4.5.	E. coli	27
4.4.6.	Polarization curve and power output	28
Conclusio	ns and future perspectives	31
Bibliograp	ohy	33
	4.4. Performance composition 4.4.1. 4.4.2. 4.4.3. 4.4.4. 4.4.5. 4.4.6. Conclusio	4.3. Plant Selection. 4.4. Performance evaluation of the upgraded iHydroMET with the optim components. 4.4.1. Organics. 4.4.2. Total Phosphate. 4.4.3. Total Nitrogen. 4.4.4. pH. 4.4.5. E. coli 4.4.6. Polarization curve and power output. Conclusions and future perspectives. Bibliography.

Abstract

In recent decades, high socio-economic development has led to an increase in water consumption, which, in turn, generates large volumes of wastewater. The centralized wastewater treatment facilities are not foreseen as a sustainable solution to manage the water and wastewater resources due to associated issues like high capital investments, high maintenance costs, upgradation challenges, and operational difficulties in scarcely populated or unsuitable terrain regions. Nowadays, wastewater is increasingly regarded as the non-conventional water source to meet increasing water demands for different purposes. The processes used in the treatment plants determine the quality of treated effluent. In recent times, decentralised treatment processes capable of resource recovery are highly encouraged over the globe. In particular, integrated technologies that can help manage the wastewater at the point sources are emerging as promising alternatives to the existing approaches. In this context, our lab tested the idea of iHydroMET with the aim of developing sustainable decentralised technology. iHydroMET stands for Integrated Hydroponics Microbial Electrochemical Technology. It is based on the integration of physicochemical, biological, and bioelectrochemical processes in a single reactor system to remove different pollutants present in domestic wastewater. In a proof-of-concept design study, it worked in principle and proved its potential for wastewater treatment along with simultaneous resource recovery. As a logical continuation of the work, my thesis focussed mainly on optimising the reactor components of iHydroMET to improve the system performance.

We tested three major components; namely, i) support bed matrix, ii) cathode placement and iii) plant selection for better nutrient removal. Different granular activated charcoal (GAC) and cocopeat (CP) ratios were tested as the bed matrix component. GAC:CP in 20:80 combination showed removal of around 70% of COD, 69% ammonia, and 80% orthophosphate. The configuration of 3/4 immersed cathode in effluent showed higher closed circuit voltage (by a margin of 34±4 mV) compared to the cathode placement at the top of the bed matrix. Among different locally available plants, namely Vinca, Pothos, Peppermint, Lemongrass, and Brahmi, the Vinca plant outcomepted others in nutrient removal capabilities (total nitrogen; TN: 44 mg/L and total phosphorous; TP: 2.1 mg/L

within 48h). It offers additional benefits like perennial flowering and phytoremediation of heavy metals.

The semi-pilot scale iHydroMET with the selected reactor unit components was then evaluated for domestic wastewater treatment and resource recovery. The upgraded system achieved 79±6.7 %, 20±8.4 %, 10±2.7 % removal efficiencies for COD, TP, and TN after 3h of operation in the fed-batch tests. The removal efficiencies enhanced to 91±3.3 %, 32±1.1 %, 34±5.7 % after 24h. The maximum power densities were recorded at low levels of 1.2 mW/m² in series and 11.7 mW/m² in parallel connections. The UV treatment of the secondary effluent of iHydroMET resulted in complete disinfection. As per the standards of 'The Food and Agriculture Organization, United Nations', the treated effluent quality is fit for reuse in gardening purposes.

To summarize, the selected reactor components contributed to the improved COD removal performance by iHydroMET. However, the system could not substantially remove TN and TP and produced low electric power. It was most likely due to the lack of anoxic conditions in the lower part of the reactors. Hence, further modifications in system operation like creating a saturation zone in the lower part of the reactor units are needed to address the inefficient removal of nutrients from domestic wastewater and improve the electric output of the iHydroMET technology.

Chapter 1

Introduction

1.1 Wastewater: An environmental problem but promising solution to address water stress

Water is an essential commodity of life. All types of life formsdepend on water for survival as almost all cell functions require water directly or indirectly. Besides biological importance, water is used for a wide variety of daily activities like washing, cooking, industrial application, agriculture, etc., due to its unique chemical and physical properties. Due to rapid socio-economical development over the recent decades, water is used extensively and unethically by humans, resulting in an imbalance in the water use cycle. It has led this crucial resource, i.e., freshwater, on the verge of scarcity. One-quarter of the world's population in developing and poorly developed countries is affected by moderate to high water stress (Sands, 2020). The global population boom and urbanization expansion over the last decades also created a challenge of managing huge quantities of generated wastewater. The domestic wastewater streams, agricultural runoffs, and industrial discharges contribute massively to freshwater pollution, making freshwater unfit for direct human and livestock usage. In water-stressed countries like India, freshwater supply is projected to become an alarming issue with time. According to a 2018 Niti Aayog report, water availability is set to decline from 1,544 cubic meters per capita in 2011 (and 1,816 m³ in 2001) to 1,465 cubic meters by 2025 (NITI Aayog 2018). According to the Central Pollution Control Board, India will need 1.5 trillion cubic meters of water by 2030 (Researchers Devise New Wastewater Recycling System, India Science Wire, 2018). Hence, there is a need to manage wastewater efficiently and exploring unconventional

water sources to meet the increasing freshwater demand.

For a very long time, wastewater has been considered an environmental hazard and health concern. However, in recent years, the change in this attitude can be noticed. Now, the nonconventional sources of water, like properly treated wastewaters, are being seen as a potential source to meet future water demands. Transitioning of wastewater treatment plants into resource recovery facilities is considered a promising way to achieve resource sustainability (Wang et al., 2015). The trend is changing speedily towards maximizing resource recovery from wastewaters, mainly in the developed countries. However, in developing countries like India, the current focus is to balance the equation of wastewater produced and treated by increasing the treatment capacity. Also, strict norms and guidelines are being imposed to make existing wastewater treatment systems more efficient to produce treated effluents of desirable discharge quality and prevent pollution in aquatic freshwater bodies (Schellenberg et al., 2020). The treated wastewater can be reused to provide ecological benefits such as irrigation, flushing, and washing which ultimately reduces the demand for freshwater and augment the existing water supplies (Mo & Zhang, 2013).

1.2 Existing wastewater treatment infrastructure

Globally, more than 80% of wastewater gets discharged to various natural water bodies in the environment without much treatment or no treatment at all. In rich countries, around 70% of the wastewater gets treated, whereas, in low-income countries, the treatment is just 8% of the total wastewater generated (Wastewater, The Untapped Resource, United Nations Educational, Scientific and Cultural Organization, 2017). In India, considering the available infrastructure, existing wastewater treatment technology, and technical support, only around 34.74% of 75,020 MLD (million liters per day; data for 2018) wastewater generated could be treated in class-II towns (www.indianinfrastructure.com). The municipal wastewater generation load is so high that a high percentage (65%) of the wastewater gets discharged to surface water bodies without proper treatment causing significant pollution. Most of the existing wastewater treatment infrastructure is centralized across the globe. However, the centralized large conventional sewage treatment plants are challenged by changing climate conditions and cannot sustain the current and growing demand (Mercer, 2014). In addition, the aging wastewater collection networks suffer due

to several reasons such as costs intensive maintenance and repair and the associated harmful environmental impacts due to wastewater leakage from pipes (Electricity Use and Management in the Municipal Water Supply and Wastewater Industries, 2014). Moreover, most plants are difficult to upgrade with the new and resource recovery capable technologies. Thus, there should be more emphasis on finding alternative technologies that do not burden the environment and economy.

Initial progress in this sector has been mostly made through various government and private investments to meet Sustainable Development Goals (SDGs). It has created an opportunity for alternative approaches in cities and towns where centralized wastewater treatment systems are either absent or not efficient to accomplish the level of performance (Water, Sanitation & Hygiene, Bill & Melinda Gates Foundation, 2020). However, as the water-related sector is operated by conservative institutions that require decades to respond to the development of new technologies, there are fewer chances for an instant shift (Kiparsky et al., 2013). Whenever engineers propose novel ideas for making centralized treatment systems more efficient, the authorities tend to focus mostly on incremental improvements in the existing system rather than implementing new effectual technologies because of the possibility of incurring misfortune or failure (Rabaey et al., 2020).

1.3 The era of decentralization

As the existing centralized conventional approach has a limited network, inadequate functioning, and has cost restrictions to build and operate, we should not rely on them as the only way to manage wastewaters. Hence, alternative approaches that do not rely upon centralized sewer systems are gaining attention (Rabaey et al., 2020). Decentralised wastewater management approaches are considered a sustainable and cost-effective alternative as these systems treat and discharge or offer reuse of the effluent at the point sources. They have been promoted because of their low operational and maintenance requirements and small-scale investments. Their implementation also makes it feasible to treat wastewater from those areas that are not connected to centralized wastewater management networks. Such areas are sometimes scarcely populated and located at complex terrains; it is difficult to sustain a centralized sewage network. The major

advantage a decentralised wastewater treatment system offers is the ability of the consumer to control the quality of water delivered according to its usage (Rabaey et al., 2020). For example, if one needs treated water for gardening only, it does not require additional nutrients removal steps. The commercialization of these technologies has been started. For instance, Hydraloop turnkey systems are already on the market that enable greywater treatment for reuse as flushing water or washing purposes (Innovative Water Recycling System, Hydraloop, International, 2020). Additionally, many decentralised technologies are under investigation in different countries for managing wastewater at point sources.

1.4 Emerging decentralised technologies for wastewater treatment

Considerable efforts are going on developing technologies that are cost-effective, easy to implement, and capable of treating a sufficient amount of wastewater to meet the load at household or community levels. Developing a standalone new technology does not seem feasible in the near future, so more focus is drawn on technologies that are based on the integration of different types of existing processes or technologies. It allows using several advantages like a high treatment rate system capable of resource recovery and overcoming the limitations of the individual treatment processes. In recent times, most of the developments revolve around technologies that are ecologically engineered to mimic nature's ways of managing wastewaters. Generally, an engineered system consists of several different tanks interconnected in series through which wastewater flows. The wastewater is passed through diverse biota like aquatic macrophytes, aquatic plants, and filter feeders for wastewater treatment through ecological interactions and purification (Chiranjeevi et al., 2013). Constructed wetlands (CWs) is one such technology that offers promising solutions for wastewater treatment and water management. They are based on phytoremediation principles. In the last decade, integration of CWs with microbial fuel cells (MFC) has become a promising approach (Yadav et al., 2012). MFCs are an emerging bio-electrochemical technology for simultaneous wastewater treatment and electricity production. In recent times, further advancement in CWs has been done by replacing the non-conductive bed matrix with conductive materials. It is termed electro-wetlands (Ramírez-Vargas et al., 2018). These technologies seem to be complex but are easy to implement and manage in the real fields and consume low energy. In all cases, the main

challenge is to make technology proficient and sustainable in recovering and reusing resources like water, nutrients, and energy from the wastewater.

1.5 The idea of iHydroMET

With an aim to develop a sustainable and economically feasible solution for wastewater management at the households, the integrated drip hydroponics-MFC system design was tested for domestic wastewater treatment in our lab (Yadav et al., 2020). We refer to the technology as iHyroMET. It consisted of a drip manifold system for influent feeding and effluent collecting duct along with ten reactor units. Each unit housed three key reactor components: the cocopeat bed matrix, a graphite electrode assembly, and a lemongrass sapling. After 3h operation in a batch recirculation mode, the system achieved $72 \pm 2.4\%$ COD, $83 \pm 1.1\%$ phosphate, and $35 \pm 4.2\%$ ammonia removal efficiencies. It also yielded low levels of power output and plant biomass. The arrangement of different components reactor units facilitated various physicochemical, biological, and bio-electrochemical processes in different microenvironments, resulting in removing different pollutants in the integrated system. Overall, this low-cost and simple integration worked in principle and proved its potential for wastewater treatment along with simultaneous resource recovery. However, clogging of bed matrix and low COD and N removal were the issues faced during the operation. It should be noted that the performance achieved by the system was with unoptimised reactor components and operation parameters. Hence, the scope existed for further optimisation studies focusing on different process parameters to improve both wastewater treatment and resource recovery performance. These include operational parameters such as drip rate, feed volume and HRT, and reactor parameters such as design components, support bed materials, and plants with better growth rates and nutrient uptake capabilities. My thesis work focused on optimising the reactor design and components like bed matrix, cathode placement, and plant selection, followed by the performance evaluation of iHydroMET with the optimised components.

Chapter 2

Objectives

- To optimise the reactor components of the iHrdroMET system.
- To test the wastewater treatment performance of the upgraded iHydroMET system with optimised reactor components.

Chapter 3

Materials and Methods

3.1 Selection and optimisation of support bed matrix components

Different support bed matrices were tested to optimise a favorable support bed medium for plant growth and the proliferation of different microbial communities. The purpose was also to target the removal of both emerging pollutants and nitrogen efficiently without compromising the removal of other wastewater contaminants. These include granular activated charcoal (GAC), quartz/silica sand (QS), cocopeat (CP), and their combinations in different proportions (Photograph 1).

3.1.1 Cocopeat

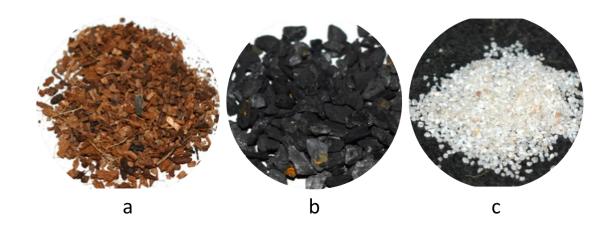
CP is known for its potential as a good support material not only to facilitate the growth of diverse microorganisms but also for anchoring plants (Awang et al., 2010). It is a natural fibre, which is extracted as a by-product from the outer coconut husk. It is widely used in horticulture and mixed with soil, which helps trap moisture for a longer time. It is very stable in terms of its physical and chemical properties. CP used in this study was manufactured in Karnataka, India, and bought from TrustBasket online store. The product was washed two times and dried at room temperature to remove its natural reddish colour and to wash off unwanted soil particles and dust before use. After drying, it was sieved to get particles of size 3-5 mm to get rid of fine-sized particles, which may cause severe clogging in the system.

3.1.2 Granular activated charcoal

It is a form of carbon derived by processing charcoal to achieve low-volume pores that increase its surface area for adsorption. It is widely used in wastewater treatment systems due to its adsorption capability of various contaminants (Hung et al., 2005). The GAC (Andel Marketing & Services) of size around 5 mm was used. It was washed twice and dried at room temperature to remove dirt and fine black charcoal particles before use.

3.1.3 Quartz sand

Quartz sand (QS) is white-yellowish sand composed of silicon dioxide. It is one of the low-cost bed materials used in conventional bio-filtration technologies like trickling filters, constructed wetlands, etc. (Dacewicz & Chmielowski, 2018). The QS of size around 2-3 mm was obtained from the Institute's STP. Before use, it was also washed and dried to remove dirt.



Photograph 1. Different materials tested as the bed matrix components: a) cocopeat, b) granular activated charcoal, and c) quartz sand

3.1.4 Bed matrix optimisation

The custom-made reactor units of dimensions 20 cm length and 7.5 cm diameter were filled up to 15 cm height with different support bed materials mentioned above, and their combinations (proportion by the length of the reactor and represented from bottom to top).

These include GAC:CP (50:50), GAC:QS (50:50), GAC:CP:QS (50:25:25), and GAC:QS:CP (50:25:25) (Fig. 1). GAC was used in all combinations. The removal of the major wastewater pollutants, namely, organic matter, phosphates, and ammonia, was then investigated by passing the wastewater through different bed matrices. All experiments were conducted in duplicate. During the acclimatization period, all the reactor units placed in a system were fed parallelly by continuously recirculating fresh 10L domestic wastewater twice a day. After attaining the stabilized performance, the domestic wastewater was fed in a down-flow manner to the individual reactor units with a flow rate of 40 ml/min with the help of a peristaltic pump for 20 min. The effluent samples were collected from the bottom of each reactor unit. The performance was assessed by analyzing different conventional wastewater parameters such as chemical oxygen demand (COD), orthophosphates, and ammoniacal-N at least three times for each reactor unit. Based on the results of different combinations of materials, we then tested different proportions of GAC and CP, such as GAC:CP (20:80), GAC:CP (40:60), and GAC:CP (60:40) in batch experiments under similar conditions (Fig. 1).

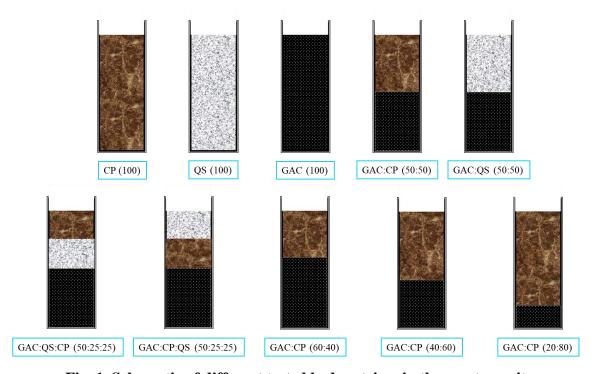


Fig. 1. Schematic of different tested bed matrices in the reactor unit

3.2 Cathode position optimisation

Graphite flat plates were used as the electrodes (IPGI Instruments, Chennai). Graphite possesses good electrical conductivity and biocompatibility and is non-corrosive. Also, it is cheaper than other types of electrode materials such as stainless steel, platinum, etc., and is readily available. Non-catalysed, rectangular-shaped graphite cathode with 4 cm x 4 cm and anodes with 2 cm x 3 cm dimension were used. Electrodes were pre-treated by acidalkali method. Aluminum connecting wires were attached to the electrodes by drilling and sealing them using conductive cement. The anode was placed vertically in the activated charcoal bed.

Three identical bed matrix reactors, housing similar graphite anodes (surface area - 6 cm²) at the bottom and cathodes (surface area - 16 cm²), were used for the experiments. The reactors were placed in effluent chambers, each filled with 500ml of fresh domestic wastewaters. The wastewater was recirculated in each reactor using a peristaltic pump at a flow rate of 50 ml/min. The bottom of the reactor units was submerged by 1 cm in the effluent chamber containing wastewater to achieve a continuous medium for closing the circuit (Fig. 2). A 1000 Ω external resistor was used for the experiment to measure the closed-circuit voltage (CCV) using a digital multimeter (Fluke 179) in both configurations. It was found to be the optimum resistance during polarisation recorded at stable voltage by varying resistance method from 50K-250 Ω using resistance box. The cathode was switched between the two tested configurations, i.e., (i) top of the reactor unit and (ii) ¾ immersed in the effluent chamber while recording the CCV data. The ¾ immersion of cathode was selected based on the report suggesting it to be the best configuration for electricity harvest (Walter et al., 2019). The readings were taken for 6h at a regular interval of 1h and were repeated three times with fresh domestic wastewater each time.

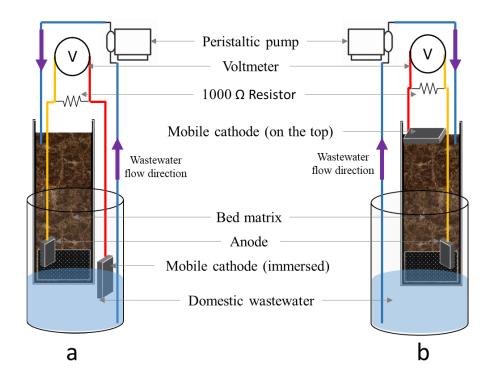
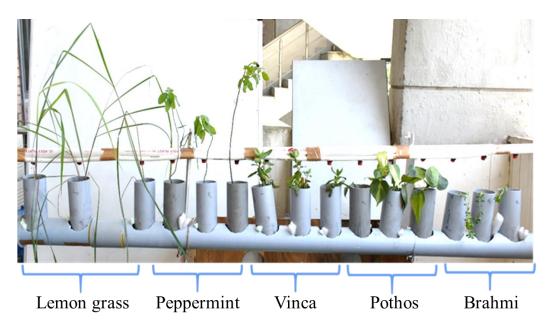


Fig. 2. Schematic of the experimental setup for harvesting electricity in two cathode positions: a) ¾ immersed in the effluent chamber and b) top of the reactor unit

3.3 Plant selection

Plants are an integral component of the integrated system. They important role in optimal wastewater treatment performance, particularly in removing nutrients N and P. Several types of plants with fibrous roots were tested for their capabilities to sustain growth in the reactor conditions. These include floricultural or aesthetic value plants (e.g., Vinca and Pothos) and commercially important plants (e.g., Peppermint, Lemongrass, and Brahmi). These were obtained from the IISER Mohali campus horticulture facility and Ajith Nursery Mohali. The experiments with different plants were conducted in triplicates. Different plants were fed with fresh domestic wastewater daily in a drip-hydroponics recirculation operation mode for a month-long period (Photograph 2). They were analyzed for easy adaptation to the domestic sewage condition, wastewater feeding operation, and visual growth. In the second set of experiments, nutrient removal by individual reactor units containing well-adapted and survived plants (i.e., Vinca and Pothos) was assessed. The

plants were fed with 500 ml of fresh domestic wastewater in a recirculation mode (Fig. 3). The samples were taken after 12h, 24h, and 48h during recirculation and were analysed for TN and TP removal. The analysis was done three times with fresh feed each time.



Photograph 2. Different test plants planted in domestic wastewater-fed reactor units

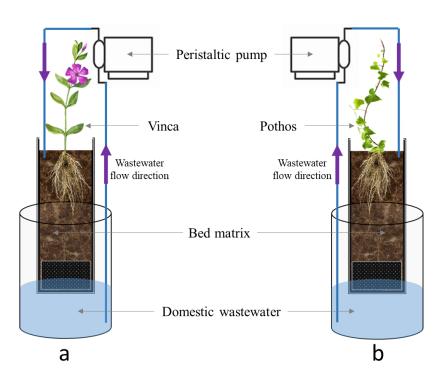


Fig. 3. Schematic of the experimental setup used to evaluate nutrient removal performance of the test plants: a) Vinca and b) Pothos

3.4 Performance evaluation of the upgraded iHydroMET system with the optimised set of reactor components

The upgraded iHydroMET system was fabricated using raw materials purchased from the local market. The system consisted of 10 reactor units, each having length of 20 cm (Figs. 4 and 5). The integrated condition comprises bed matrix (BM), electrodes-assembly, and plant sapling. All the reactor units were filled from bottom to top with, as per the selected combination (Section 3.1), i.e., GAC 20:80 CP proportion by the length of the reactor unit. The requisite number of similar stem cuttings of Vinca were propagated by placing them in tap water to develop roots and finally used for the main experiments. The graphite anode (surface area: 6 cm²) was placed within the bed matrix embedded in the GAC. The graphite cathode (surface area: 16 cm²) was vertically placed in ¾ immerged condition outside the reactor in the effluent collection duct, adjacent to each reactor unit (Fig. 4). Daily 10 L of fresh domestic sewage (DS) was collected at around 10 am from the sedimentation tank (primary treatment unit) of the sewage treatment plant facility located at IISER Mohali residential campus. In each experimental run, 10 L clarified domestic wastewater (i.e., primary effluent) was fed to the system via the drip manifold pipe. Each reactor unit received the feed at a flow rate of 50 ± 10 ml/min. The system was operated indoors under ambient temperature (25 \pm 4°C) and light illuminance (~5000 lux with 16h to 8h day and night period cycles) conditions and in a batch-recirculation drip-hydroponics mode for the wastewater feeding. Before starting the sampling period, the system was acclimatized by feeding the fresh wastewater for 15 days. The influent and the effluent samples of different hydraulic retention times (HRTs) such as 3h, 6h, 12h, and 24h were analyzed for five cycles with fresh domestic wastewater each time. The secondary effluent collected after 24h was passed through the UV system (GL6, Goodlife UV disinfection system) at a flow rate of 400ml/min. The wastewater treatment performance of the integrated system was assessed by monitoring pH (Oakton PC2700 pH meter), Chemical Oxygen Demand (COD), Biological Oxygen Demand (BOD₅), total solids, nitrogen (i.e., Total Nitrogen, ammoniacal-N, nitrate-N, and nitrite-N), total phosphates, and MPN test for E.coli.

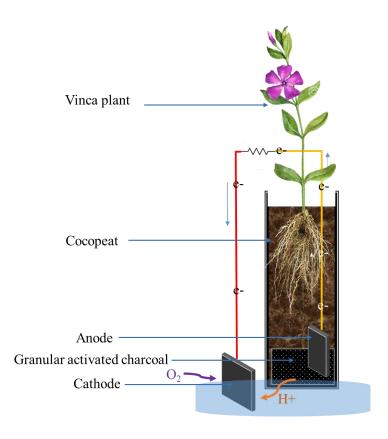


Fig. 4. Schematic of a single reactor unit configuration with the major components.

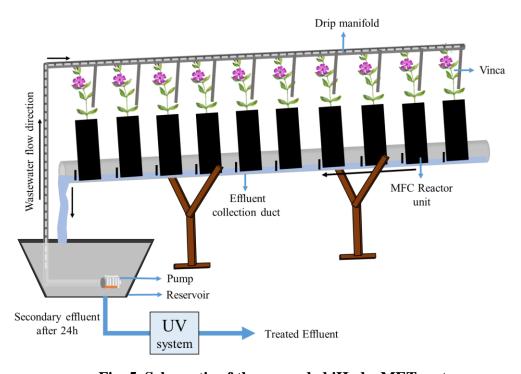


Fig. 5. Schematic of the upgraded iHydroMET system.

3.5 Analyses and calculations

3.5.1 Wastewater parameters

Different wastewater parameters were analysed according to the goal of each experiment. The pH was estimated using Oakton PC2700 pH meter, and BOD was estimated with the help of Oxitop BOD method, incubation at 20°C. The COD, ammonium-N, orthophosphate-P and E.coli were estimated by closed reflux, phenate, stannous chloride colorimetric and MPN methods, respectively using standard protocols adapted from "Standard Methods for the Examination of Water and Wastewater" (APHA, 2012). Nitrate-N and nitrite-N were estimated by reagent-free method (OptRF measurement) of UV-Vis spectrophotometer (Photolab 7600, Xylem Analytics, Germany). TN and TP concentrations were analyzed by the alkaline persulfate digestion method (De Borba et al., 2014). For TN, 2 ml sample was taken in digestion vials and 0.3 ml of alkaline persulfate digestion solution (0.3 gm of potassium persulfate in 5 ml of 1.5M NaOH solution) was added to it. The samples were digested at 120°C for 30 min and then analysed and quantified using UV-Vis spectrophotometer by taking the spectrum for the range of 220-230 nm and taking the double derivative of it. For TP, 2.5 ml of sample was taken in digestion vials, and 50 µL of H₂SO₄ solution was added to it. Then 20 mg of ammonium persulfate was added to the samples. The vials were then put for digestion at 120°C for 60 min. These samples were analysed for orthophosphate using a colorimetric test.

Standard calibration curves for COD, TN, TP, and ammonia at different concentrations were plotted using spectrophotometry (Photolab 7600 UV-VIS) and used to estimate the concentrations of these constituents in the test samples. The treatment performance was evaluated by estimating removal efficiencies (ξ) calculated using Eq. (1), where $C_{\rm O}$ represents the initial concentration of the specific pollutant in the influent, and $C_{\rm S}$ denotes the concentration of the same pollutant (mg/l) in the effluent.

$$\xi = \frac{C_o - C_s}{C_o} \times 100$$
 Eq. (1)

The data are presented as averages along with uncertainties based on at least three analyses for each parameter in all experiments. All the visual observations were recorded in pictorial

form using a Nikon D3400 DSLR camera.

3.5.2 Electrochemical parameters

Polarization behavior was recorded during the stable voltage phase of operation. The reactors of iHydroMET were connected in series and parallel configuration to assess energy harvest in a closed circuit mode. The polarization test was performed to determine the cell design point of the system using a various range of resistances in both series ($50k-100~\Omega$) and parallel ($1000-5~\Omega$) configurations. The closed-circuit voltage was recorded using the auto-range digital multimeter and datalogger (Keithley 2400). The CCV at the determined optimum resistance was also recorded for 6h in two cycles, each starting with a fresh load, for both series and parallel connections.

The slope of the polarization curve represents the internal resistance of the system, which was calculated using Eq. (2)

$$R_{int} = -\frac{dE}{dI}$$
 Eq. (2)

Where.

 R_{int} : internal resistance (Ω),

E: potential (V)

I: current (A)

During the closed-circuit mode operation, the current (I) was calculated using Ohm's law, i.e., V = IR, and Power (mW) was derived from P=IV equation. Current density (mA/m²) and power density (mW/m²) were calculated by normalizing the respective absolute current and power values with the cathode surface area (0.0016 m²).

Chapter 4

Results and Discussion

4.1 Selection and optimisation of support bed matrix components

All the tested bed material components and their combinations in the reactor units achieved stabilized performance within 10-15 days of wastewater feeding. The CP was the best performing support bed material among all for COD, ammonia, and phosphate removal (Table 1). These results validated the use of CP in our earlier experiments and suggested its use in future experiments as well. It has a high water-holding capacity which helps to stop the oxygen diffusion in the medium and helps to achieve anaerobic conditions. It supports plant growth and microbial growth by giving support and a nutrient-rich environment due to its high absorption capabilities (Awang et al., 2009). The data suggest that CP facilitated microbial degradation of organic matter and ammoniacal nitrogen conversion and phosphate adsorption more efficiently than the QS and GAC materials. QS was not effective in the removal of pollutants. It is primarily used as a filter bed without much associated physicochemical, specifically, adsorption or biological activities. GAC also did not show significant removal of COD, ammonia, and phosphate as compared to CP. However, it has been reported as good material for facilitating nitrate removal in the anoxic environment via microbial denitrification and as a high surface area adsorbent for emerging micropollutants like pharmaceuticals and personal care products (PPCPs) from the wastewater (Elumalai et al., 2014; Ohe et al., 2003). One gram of activated charcoal can have a surface area of up to 3,000 m² due to this high microporosity (Saleem et al., 2019). Moreover, the inclusion of conductive material like GAC at the bottom of the bed matrix can act as the additional electron acceptor for electroactive microbes, contributing to more electricity harvest (Caizán-Juanarena et al., 2020). Hence, we tested different proportions of GAC and CP, such as GAC:CP (20:80), GAC:CP (40:60), and GAC:CP (60:40). Among these three combinations, GAC:CP (20:80) combination showed comparable removal efficiencies to the only CP bed material (Table 1). Hence, it was selected for further experiments. The GAC offers additional benefits of nitrate and emerging pollutants (like PPCPs) removal (Ohe et al., 2003).

Table 1. COD, ammoniacal-N, and orthophosphate removal efficiencies (%) achieved by different support bed material matrices in reactor units.

S. No.	Bed matrix	Removal efficiencies (%)			
	components	COD	NH ₃ -N	Orthophosphate	
	(proportion %)				
1	CP (100)	71.80 ± 2.64	79.80 ± 4.28	80.61 ± 1.18	
2	QS (100)	65.11 ± 3.47	66.67 ± 6.69	73.23 ± 5.53	
3	GAC (100)	35.42 ± 2.71	37.35 ± 3.49	66.24 ± 4.12	
4	GAC:CP (50:50)	57.79 ± 2.67	58.16 ± 2.03	74.17 ± 1.51	
5	GAC:QS (50:50)	50.60 ± 2.96	54.78 ± 1.90	70.52 ± 2.4	
6	GAC:CP:QS	48.44 ± 3.66	59.05 ± 1.55	72.03 ± 4.71	
	(50:25:25)				
7	GAC:QS:CP	45.31 ± 1.45	54.75 ± 2.03	69.82 ± 1.65	
	(50:25:25)				
8	GAC:CP (20:80)	69.69 ± 1.94	68.75 ± 2.53	79.67 ± 1.33	
9	GAC:CP (40:60)	65.22 ± 0.65	61.40 ± 2.31	77.32 ± 2.12	
10	GAC:CP (60:40)	52.12 ± 2.45	56.23 ± 1.93	71.38 ± 3.46	

4.2 Cathode position optimistion

The position of the electrodes in the reactors plays a crucial role in electricity harvest. Cathodes are usually reported as a limiting component for the oxygen reduction reaction, hence tested for proper placement in the reactor units (Khotseng, 2018). The CCV recorded

at an optimised external load of $1000~\Omega$, in the two tested cathode placement conditions i.e., a) on the top of bed matrix and b) $\frac{3}{4}$ immersed in the effluent (Walter et al., 2019), were in the range of $47.17 \pm 21.21~\text{mV}$ to $54.67 \pm 24.32~\text{mV}$ and $77.75 \pm 31.25~\text{mV}$ to $92.41 \pm 32~\text{mV}$ respectively during 6h of operation. The maximum CCV was recorded after 1h of operation, and thereafter it gradually decreased over the operation time irrespective of the tested condition (Fig. 6). It might be because the organic load was initially higher, and it decreased slowly with time. The CCV obtained in the case of cathode immersed in the effluent was higher (by a margin of $34\pm4\text{mV}$) compared to cathode placed on the top of the bed matrix. Three possible reasons for higher CCV in the immersed cathode condition as compared to top placed cathode include (i) short effective distance (1-2 cm) between the electrodes in the former condition as compared to the later (13-14 cm) (Sajana & Mitra, 2013), (ii) the liquid medium condition between electrodes in former as compared to the less conductive solid bed matrix in the later condition, and (iii) direction of the hydrogen ion movement according to the wastewater flow making available more H⁺ ions for the oxygen reduction reaction at the cathode in the former case.

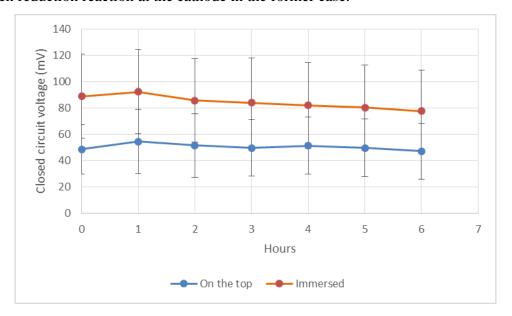


Fig. 6. Closed-circuit voltage in two cathode placement conditions for 6h.

4.3 Plant Selection

Domestic wastewater has a high level of nutrient concentration in terms of nitrogen and phosphorus. All plants require these essential nutrients for their growth and development,

and they uptake them from the surroundings (Roy et al., 2006). Among the tested plants, Brahmi and lemongrass died due to unfavorable nutrient conditions or system feeding pattern and external factors like pest attack on lemongrass. The rest of the three plants, i.e., Vinca, Pothos, and Peppermint, survived throughout the experimental cycle, but Peppermint did not show substantial growth (Fig. 7). Vinca and Pothos showed fast growth. The initial concentration of TN was 52±11 mg/L and TP was 4.2±0.5 mg/L in the fresh domestic wastewater. The final TN concentration after 48h was around 18 mg/L for Pothos and 8 mg/L for Vinca (Fig. 8). The final TP concentration after 48 hours was 2.7 mg/L for Pothos and 2.1 mg/L for Vinca (Fig. 9). The removal of TN (Vinca: 44 mg/L; Pothos: 34 mg/L) and TP (Vinca: 2.1 mg/L; Pothos: 1.5 mg/L) was slightly higher for Vinca compared to Pothos in 48h tests. These observations are in agreement with reported studies on Vinca's capability for nutrient removal from wastewater and phytoremediation of heavy metals like cadmium, lead, zinc, copper, and nickel (Frantz, 2013; Barbosa et al., 2020; Pandey et al., 2007). Based on better nutrient removal results and considering Vinca's additional benefit of producing attractive flowers compared to Pothos, it was selected as the potential plant candidate for the iHydroMET technology.

	Vinca	Pothos	Peppermint	Brahmi	Lemon grass
Before	v _i			B ₁	
After 1 month					

Fig. 7. Pictorial view showing the increment in shoot biomass of representative tested plants grown for 1 month: a) Vinca; b) Pothos; c) Peppermint; d) Brahmi, and e)

Lemongrass

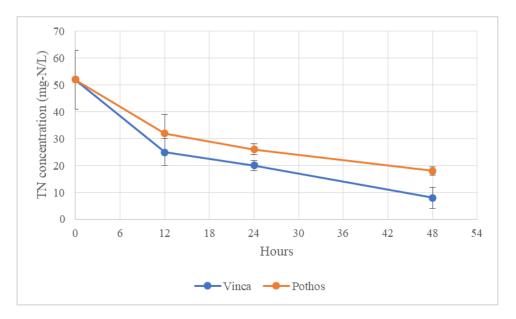


Fig. 8. Removal of total nitrogen (TN) by the tested plants for 48h period

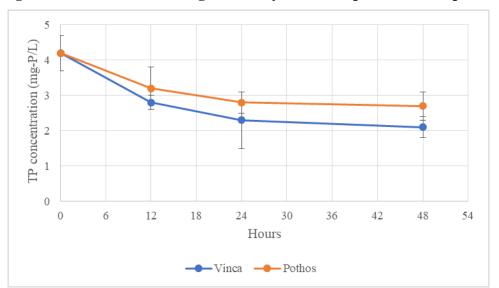


Fig. 9. Removal of total phosphorus (TP) by the tested plants for 48h period

4.4 Performance evaluation of the upgraded iHydroMET with the optimised reactor components

The characteristics of the influent (primary clarified) domestic wastewater and effluent are shown in Table 2. The wastewater treatment performance of the upgraded iHydroMET in terms of removal efficiencies of different components achieved at different HRTs is discussed in the following subsections.

 $\begin{tabular}{ll} Table 2. Consolidated data of wastewater treatment performance parameters of \\ iHydroMET at different HRTs \end{tabular}$

		Effluent at different HRTs			
Parameters	Influent	3h	6h	12h	24h
pН	7.62±0.14	7.68±0.03	7.79±0.01	7.90±0.02	7.95±0.01
Conductivity	883±84.85	837±4.82	798.5±1	767.5±1.76	749±3.6
(µS/cm)					
COD (mg/L);	237±66	48±16;	35±16;	23±13;	22±8;
(%)		(79±6.7)	(84±6.3)	(90±3.1)	(91±3.3)
BOD (mg/L);	193±46	40±7	30±1	8±4	7±3
(%)		(79)	(84)	(96)	(97)
BOD:COD	0.81±0.09	0.71±0.15	0.68±0.1	0.31±0.07	0.27±0.12
Total	6.59±0.36	5.26±0.27;	4.65±0.22;	4.59±0.19;	4.76±0.19;
Phosphorus		(20 ± 8.4)	(28±0.6)	(29±0.9)	(32±1.1)
(mg/L); (%)					
Ammonia	30.78±5.2	8.99±2.23;	1.37±3.41;	0.09±0.65;	0±0.07;
(mg/L); (%)		(70 ± 2.2)	(95±3.4)	(99±0.6)	(100)
Nitrate (mg/L)	1.49±0.26	17.78±0.44	25.44±0.42	32.9±1.92	36.27±2.42
Nitrite (mg/L)	0.56±0.49	3.81±0.2	5.81±0.82	0.48±0.83	1.12±1.12
Total Nitrogen	38.14±7.45	33.96±3.93;	31.89±4.46;	28.63±7.19;	27.89±8.48;
(mg/L); (%)		(10±2.7)	(16±9.2)	(25±7.4)	(34±5.7)

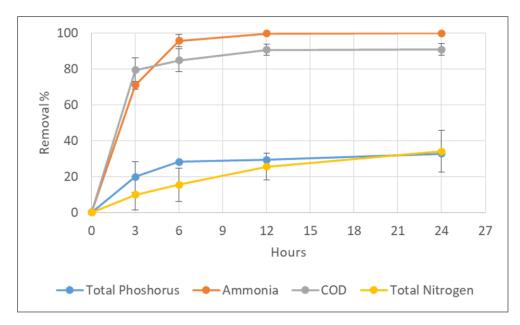


Fig. 10. Removal efficiencies of various pollutants by iHydroMET at different HRTs

4.4.1 Organics

Domestic wastewater is mainly composed of a biodegradable fraction of organics. The average BOD:COD ratio of the influent was 0.81 ± 0.1 during the experimental period. It suggests that a majority of the organics content is biodegradable (Choi et al., 2017). Hence, the BOD:COD ratio decreased with time as the fraction of biodegradable organics was consumed/degraded by the microbes (Table 2). About 80% of the organic matter was removed in the first 3h of operation due to entrapment of the organics in the bed matrix, which can be further degraded by microorganisms. Around 90% COD removal efficiency was achieved after 24h of operation (Fig. 10). It is mainly due to the following reasons: (i) the increase in the bed matrix component, i.e., cocopeat and GAC. Due to an increase in height, the wastewater had to travel longer in the reactor, which leads to maximum adsorption of the organics in the bed matrix and eventual degradation by microbes. (ii) The wastewater is recirculated in the reactor, which provides sufficient aeration for aerobic microbial interactions to oxidise organic matter (Tantak et al., 2014). (iii) In addition to aerobic and anoxic microbial activities, the diverse microbiota in the rhizospheric zone can also contribute to COD removal by facilitating conversion of complex organic forms into simpler ones that can be used up by other microbes. (iv) The electrogenic microorganisms present in the anodic zone oxidise organic matter using the anode as an electron acceptor and further enhance organics removal (Jiang et al., 2021). Overall, the upgraded system showed a significant improvement, from 72% removal (in the earlier system) to 80%, in terms of COD removal after 3h of operation (Yadav et al., 2020).

4.4.2 Total Phosphate

The maximum phosphate removal efficiency of around 32% was achieved in the upgraded system after 24h operation. Plant uptake and adsorption of orthophosphate on bed matrix are the two key activities contributing to TP removal. From Fig. 10, it is clear that most of the removal was observed in the first 6h, which might be due to the trapping and adsorption (Shi et al., 2011), along with plant uptake of different types of phosphate forms (Frantz, 2013). However, no removal occurred after 6h suggesting a low requirement of phosphorus by plants. The specific microbes, i.e., phosphate-accumulating microorganisms (PAOs) are also reported for phosphate removal under controlled alternate aerobic and anaerobic conditions (Ubukata, 2006). However, it was most likely not the case in the tested reactor units due to the inadequate anaerobic zone.

4.4.3 Total Nitrogen

The characterization of influent shows higher concentration of ammonia (31 mg/L) and lower nitrite (0.6 mg/L) and nitrate (1.5 mg/L) concentrations. After 6h of operation, almost all ammonia got converted to other nitrogen forms, i.e., nitrite (5.8 mg/L) and nitrate (25.4 mg/L) due to microbial nitrification processes facilitated by continuous aeration and mixing of wastewater due to recirculation. We earlier hypothesized that the presence of cocopeat might facilitate a significant anaerobic microbial environment in the lower reactor portion, enhancing the denitrification process and conversion of nitrite to nitrate and further nitrate reduction to free nitrogen. However, from results, it is clear that the effect of the denitrification process was not so prominent, resulting in the accumulation of nitrate and nitrite forms in the effluent. It suggests that the wastewater feeding mode and process were not appropriate to achieve a prominent anoxic zone for better removal of nitrate and nitrite by the denitrification process. In addition, the continuous recirculation of aerated wastewater does not allow its retention for a longer duration, thereby creating a low or

negligible anoxic environment. Overall, around 34% removal in the TN concentration was observed after 24h operation. From Fig. 10, for the first 12h there is a steady increase in the removal efficiency, which might be mostly due to the trapping and adsorption along with plant uptake of different types of nitrogen forms (Ohe et al., 2003). However, the slow removal trend after 12h can be justified by only the plant uptake process. Apart from the conversion of ammonia into nitrite and nitrite, the organic nitrogen forms get trapped or adsorbed in the bed matrix, which are then converted to simpler inorganic nitrogen forms by microbial interactions. Plants can ultimately take these up for its growth. It can be confirmed by the visual growth observed in Vinca. An emerging anammox process, which is reported for ammonia oxidation process at the anodic zone, might not be happening due to inadequate anaerobic zone in our reactor units (Qu et al., 2014).

4.4.4 pH

pH is very crucial for biological interactions, chemical and physical treatment of the wastewater. Biological processes would not function properly at abnormal pH, leading to a decrease in the pollutant removal rates. pH was monitored throughout the operation cycle, which co-relates with the stable removal efficiencies. Influent pH was 7.6. As the time progressed, the pH increased (3h - 7.7; 6h - 7.8; 12h - 7.9; 24h - 8.0). It might be happening due to the stripping of CO₂ due to operational conditions and organic matter degradation (Cohen & Kirchmann, 2004).

4.4.5 E. coli

Fecal coliforms (predominantly *E. coli*), which originate from fecal wastes of humans and animals, are found in wastewater streams. The influent MPN was $(3.6\pm1.5)\times10^{11}$ MPN/100ml. There was no removal of coliforms after 24h operation in the secondary effluent. It suggests that the iHydroMET system is not efficient for the removal of microbial load. Hence, the UV treatment system was used. It achieved the removal of almost all coliforms since the final treated effluent showed a negligible count (<2 MPN index/100 ml) for *E. coli*. The disinfected iHydroMET effluent can be used for different reuse purposes.

4.4.6 Polarization curve and power output

Polarization behavior is used for analyzing the electrical output performance of the fuel cells. With an increase in resistance, there was a decrease in current. At low resistance, current density was high, but the potential was low because more electrons flow in the circuit at low resistance. At maximum load, 50 k Ω in series and 1000 Ω in parallel, negligible current output was recorded. At the resistances of 9000 Ω in series and 150 Ω in parallel, maximum power densities were observed, i.e., 1.2 mW/m² in series and 11.7 mW/m² in parallel connection (Fig. 11). In the proof of principle study, power densities were 30.9 mW/m² and 31.6 mW/m² in series and parallel connections, respectively (Yadav et al., 2020). The lower power density output might have happened in the upgraded system because the organic load in the influent was low as COD in the influent was around 200 mg/L less compared to the proof of principle study (COD: 411±32 mg/L). It is reported that the power output is directly dependent on the organic load of the wastewater (Velvizhi & Venkata Mohan, 2012). In addition, most of the complex organics might have been trapped in the upper bed matrix of the reactor unit, which can be likely degraded by the aerobic microbes without contributing to energy harvest. Some loss of organics might also be happening while reaching the bottom zone of the reactor unit, covering a distance of around 10 cm. Hence, at the active anodic bottom zone, a low amount of organics would be available for electroactive microbes to produce free electrons and hydrogen ions. Also, due to poor anoxic conditions in the anodic region, fewer microbes can use anode as the terminal electron acceptor as aerobic or facultative microbes dominate such environments.

In serially connected MFCs, the output voltage will be higher, but the current will be less, whereas in parallelly connected MFCs, the output current will be higher, but the voltage will be less. As per the appliance requirement, the type of connection, i.e., serial or parallel, can be chosen. The combination of series and parallel connection of MFCs can be done to optimise the energy harvest of the whole system.

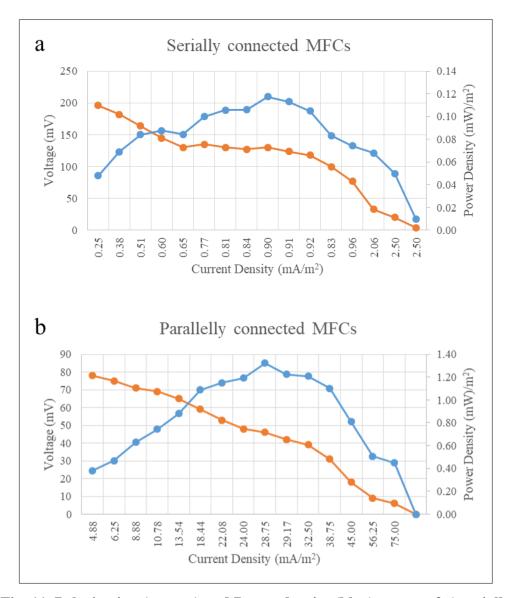
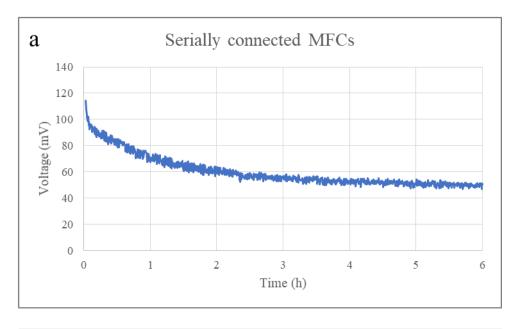


Fig. 11. Polarization (orange) and Power density (blue) curves of a) serially connected MFCs, and b) parallelly connected MFCs of the iHydroMET system.

At the determined resistances (9k Ω and 150 Ω in series and parallel, respectively), the long-term closed-circuit power generation performance was recorded for 6h (Fig. 12). The CCV dropped with time and stabilised at ~50 mV and ~10 mV in series and parallel connections, respectively. It is the actual energy harvest during the system operation.



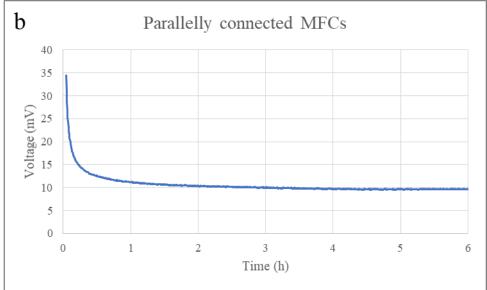


Fig. 12. Closed-circuit potential curves for 6h at optimum resistance with the a) serially connected MFCs, and b) parallelly connected MFCs of the iHdroMET system.

Chapter 5

Conclusions and future perspectives

In this work, the optimisation of different reactor components for the iHydroMET system was successfully achieved. The combination of GAC:CP (20:80) was found to be the best performing bed matrix to facilitate the removal of nitrogen, organics, and phosphates. The cathode placement in the effluent was the best for maximum energy harvest in the tested reactor units. Among different plants, Vinca was chosen for further studies because of its efficient capabilities for removing nutrients such as N & P and improving the aesthetics of the system.

The upgraded iHydroMET with the optimised reactor components showed efficient organic removal (up to 90% within 24h) but could not achieve the TN and TP removal efficiencies to the desired levels. UV disinfection further enhanced the effluent quality of iHydroMET. The effluent can be used for different purposes like gardening and flushing, as the effluent contains a good percentage of N and P, which is free of harmful microbes (Pescod, 1992). TP removal was less due to the low phosphorus requirement by plants and inefficient PAOs' activity because of the inadequate anoxic zone. TN removal was low because the denitrification process was not prominent due to the absence of anoxic conditions and continuous recirculation of aerated wastewater. The system may be modified slightly by submerging the reactor units half in the wastewater to make a significant anoxic zone and enable the denitrification process in the lower portion of the reactor, which was the limiting process in this study.

The TN removal efficiencies might be improved considerably by creating a saturation zone

at the bottom region of the reactor unit. This modification may also increase the system's power output due to favorable growth conditions for electroactive microorganisms. In addition, the combination of series and parallel electrical connections of reactor units can be optimised to get maximum energy harvest from the whole system. The TP removal can be enhanced by maintaining controlled alternate aerobic and anaerobic conditions to enrich PAOs. Also, the microbial community assessment can be investigated to improve performance further. Recirculation can also be avoided to stop continuous mixing of treated effluent to the influent, which is not an effective way of feeding. A combination of serially and parallelly hydraulically connected reactor unit combinations can be tried. The optimisation studies focusing on different parameters such as mode of operation, hydraulic retention time (HRT) or recirculation time, drip rate to each reactor unit, hydraulic and organic loading rate need to be conducted to enhance the performance of iHydroMET. In conclusion, my thesis work suggests that a better performing iHydroMET can be configured with further improvements, which can be implemented as a low-cost treatment technology for managing wastewater at the points sources.

Bibliography

- 2018 NITI Aayog. (2018). Retrieved April 6, 2021, from https://niti.gov.in/
- 2017 -Wastewater, The Untapped Resource / United Nations Educational, Scientific and Cultural Organization. (2017). Retrieved April 6, 2021, from http://www.unesco.org/new/en/natural
 - sciences/environment/water/wwap/wwdr/2017-wastewater-the-untapped-resource/
- APHA Method 9221: Standard Methods for the Examination of Water and Wastewater. (2012). 13.
- Awang, Y., Shaharom, A. S., Mohamad, R. B., & Selamat, A. (2009). Chemical and Physical Characteristics of Cocopeat-Based Media Mixtures and Their Effects on the Growth and Development of *Celosia cristata*. *American Journal of Agricultural and Biological Sciences*, 4(1), 63–71. https://doi.org/10.3844/ajabssp.2009.63.71
- Awang, Y., Shaharom, A. S., Mohamad, R. B., & Selamat, A. (2010). Growth Dynamics of *Celosia cristata* Grown in Cocopeat, Burnt Rice Hull and Kenaf Core Fiber Mixtures. *American Journal of Agricultural and Biological Sciences*, *5*(1), 70–76. https://doi.org/10.3844/ajabssp.2010.70.76
- Barbosa, É. S., Cacique, A. P., de Pinho, G. P., & Silvério, F. O. (2020). Catharanthus roseus potential for phyto-stabilizing metals in sewage sludge. *Journal of Environmental Science and Health. Part A, Toxic/Hazardous Substances & Environmental Engineering*, 55(3), 209–215. https://doi.org/10.1080/10934529.2019.1680059
- Caizán-Juanarena, L., Sleutels, T., Borsje, C., & ter Heijne, A. (2020). Considerations for application of granular activated carbon as capacitive bioanode in bioelectrochemical systems. *Renewable Energy*, 157, 782–792. https://doi.org/10.1016/j.renene.2020.05.049
- Chiranjeevi, P., Chandra, R., & Mohan, S. V. (2013). Ecologically engineered submerged and emergent macrophyte based system: An integrated eco-electrogenic design for harnessing power with simultaneous wastewater treatment. *Ecological Engineering*, *51*, 181–190. https://doi.org/10.1016/j.ecoleng.2012.12.014
- Choi, Y.-Y., Baek, S.-R., Kim, J.-I., Choi, J.-W., Hur, J., Lee, T.-U., Park, C.-J., & Lee, B.

- J. (2017). Characteristics and Biodegradability of Wastewater Organic Matter in Municipal Wastewater Treatment Plants Collecting Domestic Wastewater and Industrial Discharge. *Water*, *9*(6), 409. https://doi.org/10.3390/w9060409
- Cohen, Y., & Kirchmann, H. (2004). Increasing the pH of Wastewater to High Levels with Different Gases—CO ₂ Stripping. *Water, Air, & Soil Pollution*, *159*(1), 265–275. https://doi.org/10.1023/B:WATE.0000049185.69759.b0
- Dacewicz, E., & Chmielowski, K. (2018). The Importance of Media in Wastewater Treatment. In I. X. Zhu (Ed.), *Sewage*. InTech. https://doi.org/10.5772/intechopen.75625
- De Borba, B. M., Jack, R. F., Rohrer, J. S., Wirt, J., & Wang, D. (2014). Simultaneous determination of total nitrogen and total phosphorus in environmental waters using alkaline persulfate digestion and ion chromatography. *Journal of Chromatography A*, *1369*, 131–137. https://doi.org/10.1016/j.chroma.2014.10.027
- Electricity Use and Management in the Municipal Water Supply and Wastewater Industries. (2014). 194.
- Elumalai, E. K., Kayalvizhi, K., & Silvan, S. (2014). Coconut water assisted green synthesis of silver nanoparticles. *Journal of Pharmacy & Bioallied Sciences*, 6(4), 241–245. https://doi.org/10.4103/0975-7406.142953
- Frantz, J. M. (2013). Uptake Efficiency of Phosphorus in Different Light Environments by Zinnia (Zinnia elegans) and Vinca (Catharanthus roseus). *HortScience*, 48(5), 594–600. https://doi.org/10.21273/HORTSCI.48.5.594
- Hung, Y.-T., Lo, H. H., Wang, L. K., Taricska, J. R., & Li, K. H. (2005). Granular Activated
 Carbon Adsorption. In L. K. Wang, Y.-T. Hung, & N. K. Shammas (Eds.),
 Physicochemical Treatment Processes (pp. 573–633). Humana Press.
 https://doi.org/10.1385/1-59259-820-x:573
- Indian Infrastructure Magazine / Infrastructure Magazine India. (n.d.). Indian Infrastructure. Retrieved April 6, 2021, from https://indianinfrastructure.com/
- Innovative Water Recycling System / Hydraloop / International. (2020). Retrieved April 6, 2021, from https://www.hydraloop.com/
- Jiang, Y., Zhao, H., Liang, J., Yue, L., Li, T., Luo, Y., Liu, Q., Lu, S., Asiri, A. M., Gong, Z., & Sun, X. (2021). Anodic oxidation for the degradation of organic pollutants:

- Anode materials, operating conditions and mechanisms. A mini review. *Electrochemistry Communications*, 123, 106912. https://doi.org/10.1016/j.elecom.2020.106912
- Khotseng, L. (2018). Oxygen Reduction Reaction. *Electrocatalysts for Fuel Cells and Hydrogen Evolution Theory to Design*. https://doi.org/10.5772/intechopen.79098
- Kiparsky, M., Sedlak, D. L., Thompson, B. H., & Truffer, B. (2013). The Innovation Deficit in Urban Water: The Need for an Integrated Perspective on Institutions, Organizations, and Technology. *Environmental Engineering Science*, *30*(8), 395–408. https://doi.org/10.1089/ees.2012.0427
- Mercer, K. (2014). 2014 AWWA State of the Water Industry Report. 29.
- Mo, W., & Zhang, Q. (2013). Energy–nutrients–water nexus: Integrated resource recovery in municipal wastewater treatment plants. *Journal of Environmental Management*, 127, 255–267. https://doi.org/10.1016/j.jenvman.2013.05.007
- Ohe, K., Nagae, Y., Nakamura, S., & Baba, Y. (2003). Removal of Nitrate Anion by Carbonaceous Materials Prepared from Bamboo and Coconut Shell. *Journal of Chemical Engineering of Japan*, 36(4), 511–515. https://doi.org/10.1252/jcej.36.511
- Pandey, S., Gupta, K., & Mukherjee, A. K. (2007). *Impact of cadmium and lead on Catharanthus roseus—A phytoremediation study*. 8.
- Pescod, M. B. (1992). *Wastewater Treatment and Use in Agriculture*. Food and Agriculture Organization of the United Nations.
- Puchongkawarin, C., Gomez-Mont, C., Stuckey, D. C., & Chachuat, B. (2015). Optimisation-based methodology for the development of wastewater facilities for energy and nutrient recovery. *Chemosphere*, 140, 150–158. https://doi.org/10.1016/j.chemosphere.2014.08.061
- Qu, B., Fan, B., Zhu, S., & Zheng, Y. (2014). Anaerobic ammonium oxidation with an anode as the electron acceptor. *Environmental Microbiology Reports*, 6(1), 100–105. https://doi.org/10.1111/1758-2229.12113
- Rabaey, K., Vandekerckhove, T., de Walle, A. V., & Sedlak, D. L. (2020). The third route: Using extreme decentralization to create resilient urban water systems. *Water Research*, *185*, 116276. https://doi.org/10.1016/j.watres.2020.116276

- Ramírez-Vargas, C. A., Prado, A., Arias, C. A., Carvalho, P. N., Esteve-Núñez, A., & Brix,
 H. (2018). Microbial Electrochemical Technologies for Wastewater Treatment:
 Principles and Evolution from Microbial Fuel Cells to Bioelectrochemical-Based
 Constructed Wetlands. Water, 10(9), 1128. https://doi.org/10.3390/w10091128
- Researchers devise new wastewater recycling system—India Science Wire. (2018).

 @indianscinews. Retrieved April 6, 2021, from http://vigyanprasar.gov.in/isw/Researchers-devise-new-wastewater-recycling-system.html
- Roy, R. N., Finck, A., Blair, G. J., & Tandon, H. L. S. (2006). Plant nutrition for food security. A guide for integrated nutrient management. *FAO Fertilizer and Plant Nutrition**Bulletin** (FAO). https://agris.fao.org/agris-search/search.do?recordID=XF2006427439
- Sajana, T. K., & Mitra, A. (2013). Effect of pH and distance between electrodes on the performance of a sediment microbial fuel cell. *Water Science and Technology : A Journal of the International Association on Water Pollution Research*, 68, 537–543. https://doi.org/10.2166/wst.2013.271
- Saleem, J., Shahid, U. B., Hijab, M., Mackey, H., & McKay, G. (2019). Production and applications of activated carbons as adsorbents from olive stones. *Biomass Conversion and Biorefinery*, *9*(4), 775–802. https://doi.org/10.1007/s13399-019-00473-7
- Sands, R. (2020). The Financing of Water: Putting the Trust in Conservation [The Policy Corner]. In *The Policy Corner*. https://www.policycorner.org/2020/05/31/the-financing-of-water-putting-the-trust-in-conservation/
- Schellenberg, T., Subramanian, V., Ganeshan, G., Tompkins, D., & Pradeep, R. (2020). Wastewater Discharge Standards in the Evolving Context of Urban Sustainability—
 The Case of India. *Frontiers in Environmental Science*, 8. https://doi.org/10.3389/fenvs.2020.00030
- Shi, Z., Liu, F., & Yao, S. (2011). Adsorptive removal of phosphate from aqueous solutions using activated carbon loaded with Fe(III) oxide. *New Carbon Materials*, 26(4), 299–306. https://doi.org/10.1016/S1872-5805(11)60083-8
- Tantak, N., Chandan, N., & Raina, P. (2014). Chapter 9 An Introduction to Biological

- Treatment and Successful Application of the Aqua EMBR System in Treating Effluent Generated from a Chemical Manufacturing Unit: A Case Study. In V. V. Ranade & V. M. Bhandari (Eds.), *Industrial Wastewater Treatment, Recycling and Reuse* (pp. 369–397). Butterworth-Heinemann. https://doi.org/10.1016/B978-0-08-099968-5.00009-X
- Ubukata, Y. (2006). Fundamental Mechanisms of Phosphate Removal by Anaerobic/Aerobic Activated Sludge in Treating Municipal Wastewater. *Engineering in Life Sciences*, 6(1), 51–56. https://doi.org/10.1002/elsc.200620114
- Velvizhi, G., & Venkata Mohan, S. (2012). Electrogenic activity and electron losses under increasing organic load of recalcitrant pharmaceutical wastewater. *International Journal of Hydrogen Energy*, *37*(7), 5969–5978.
- Walter, X. A., Santoro, C., Greenman, J., & Ieropoulos, I. (2019). Self-stratifying microbial fuel cell: The importance of the cathode electrode immersion height. *International Journal of Hydrogen Energy*, 44(9), 4524–4532. https://doi.org/10.1016/j.ijhydene.2018.07.033
- Wang, X., McCarty, P. L., Liu, J., Ren, N.-Q., Lee, D.-J., Yu, H.-Q., Qian, Y., & Qu, J. (2015). Probabilistic evaluation of integrating resource recovery into wastewater treatment to improve environmental sustainability. *Proceedings of the National Academy of Sciences*, 112(5), 1630–1635. https://doi.org/10.1073/pnas.1410715112
- Water, Sanitation & Hygiene | Bill & Melinda Gates Foundation. (n.d.). Retrieved April 6, 2021, from https://www.gatesfoundation.org/our-work/programs/global-growth-and-opportunity/water-sanitation-and-hygiene
- Yadav, A. K., Dash, P., Mohanty, A., Abbassi, R., & Mishra, B. K. (2012). Performance assessment of innovative constructed wetland-microbial fuel cell for electricity production and dye removal. *Ecological Engineering*, 47, 126–131. https://doi.org/10.1016/j.ecoleng.2012.06.029
- Yadav, R. K., Chiranjeevi, P., Sukrampal, & Patil, S. A. (2020). Integrated drip hydroponics-microbial fuel cell system for wastewater treatment and resource recovery. *Bioresource Technology Reports*, 9, 100392. https://doi.org/10.1016/j.biteb.2020.100392